#### Chemical Engineering Journal 262 (2015) 823-830

Contents lists available at ScienceDirect

# Chemical Engineering Journal

journal homepage: www.elsevier.com/locate/cej

# Characterization of $TiO_2$ nanoparticle suspensions in aqueous solutions and $TiO_2$ nanoparticle retention in water-saturated columns packed with glass beads



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# HIGHLIGHTS

- We investigated the size and zeta potential of TiO<sub>2</sub> NPs suspensions.
- $\bullet$  We conducted transport experiments of  $\text{TiO}_2$  NPs in columns packed with glass beads.
- Agglomeration decreases with increasing pH and decreasing NP concentration.
- A substantial retention of the anatase TiO<sub>2</sub> NPs within the column was observed.
- Mass recovery is increasing with decreasing NP size and increasing concentration.

#### ARTICLE INFO

Article history: Received 19 August 2014 Received in revised form 13 October 2014 Accepted 14 October 2014 Available online 18 October 2014

*Keywords:* Titanium oxides Nanomaterials Porous media Transport Particle size

# G R A P H I C A L A B S T R A C T



#### ABSTRACT

In this experimental work, various suspensions of titanium dioxide (TiO<sub>2</sub>) nanoparticles (NPs) were carefully characterized, and their transport in columns packed with glass beads were thoroughly investigated. All NP suspensions were prepared with rutile–anatase and anatase TiO<sub>2</sub>. Two different methods were used for the preparation of aquatic suspensions of TiO<sub>2</sub> NPs at relatively low concentrations. The average particle size and zeta potential of each suspension were estimated in order to investigate the effect of pH and sonication time on TiO<sub>2</sub> NP agglomeration. Finally, transport experiments of the various suspensions of anatase TiO<sub>2</sub> NPs were conducted. The concentration and particle size of the NPs were measured periodically at the column outlet. Also, the accumulated mass of TiO<sub>2</sub> NPs retained in the column was determined. The experimental data suggested that NP agglomeration decreases with increasing solution pH and decreasing NP concentration. Also, it was shown that a substantial percentage of the anatase TiO<sub>2</sub> NPs injected into the experimental column were retained within the column packing due to agglomeration, and that mass recovery is increasing with decreasing NP size and increasing injected concentration. © 2014 Elsevier B.V. All rights reserved.

# 1. Introduction

Engineered metal oxide nanoparticles (NPs) have received considerable attention over the past few years due to their rapid

\* Corresponding author. Fax: +30 2610 996573. E-mail address: sygouni@upatras.gr (V. Sygouni). large-scale production and frequent use in microelectronics, paints, catalysts, cosmetics, food, pharmaceutical industries [1–4], as well as in environmental remediation applications [5,6]. Large quantities of NPs are released into sewers, which eventually reach wastewater treatment facilities, or directly into surface waters and subsurface environment by numerous point and non-point sources, including wastewater reuse, agricultural uses, landfill



Nomenclature					
C C <sub>o</sub> C <sub>max</sub>	concentration of NPs in suspension, $M/L^3$ source concentration of NPs in suspension, $M/L^3$ maximum concentration of NPs in the effluent, $M/L^3$	$t_p$ $t_s$	duration of the suspension pulse, t duration of sonication period, t		
$d_c$	collector diameter, L	Greek Letters			
$d_p$	NP diameter, L	ζ	electrokinetic zeta potential, V		
$\langle d_p \rangle$	average NP diameter, L	2	1 /		
L m <sub>o</sub> M <sub>a</sub>	length of the porous medium, L zeroth absolute temporal moment, $tM/L^3$ normalized mass accumulated within the column, defined in Eq. (3), (-)	Abbrevi dH <sub>2</sub> O IEP NB	ations deionized water isoelectric point papoparticle		
$M_r$	mass recovery, defined in Eq. (1), M	TiO.	titanium dioxide		
Q	flow rate, L <sup>3</sup> /t	1102			
ta	duration of aging period, t				

leachates, and underground storage tank leakages [7–12]. The NPs have active surfaces and can interact with molecules, organic material, biocolloids, and clays [13-15]. Consequently, NPs that enter the environment undergo surface interactions, producing very stable agglomerates, which are capable of migrating over long distances [16–24]. The toxicological risks that NPs can pose to human health and the environment have not been thoroughly examined yet [25]. To assess these potential risks, the fate and transport of NPs in environmental systems should be carefully investigated [8]. However, it should be noted that the migration behavior of NPs in subsurface formations is relatively complex [8], because it is known to be affected by many factors including solution chemistry (ionic strength, pH), temperature, flow velocity, presence of surfactants, gravity effects, and existence of biofilms [17,19,26–30]. Besides the above-mentioned environmental applications, NPs are used in Enhanced Oil Recovery for altering the interfacial tension values or contact angles in two-phase (wateroil) or three-phase flow systems (gas-water-oil) [31] or applications concerning hydraulic fracturing [32]. Also, NPs are used in medicine for functionalized bone implants [33], in photovoltaic devices [34] or memory devices [35].

Titanium dioxide  $(TiO_2)$  was employed in this experimental study, because it is one of the most frequently used nanomaterials in consumer products [36,6]. The stability of  $TiO_2$  NPs in aqueous solutions is significantly affected by solution pH, surface charge, inorganic salts, and organic matter [13,21,23,24]. Also, the initial NP concentration plays an important role in agglomerate growth, especially for concentrations ranging between 0.1 and 10 mg/L [21]. Furthermore, the retention of  $TiO_2$  NPs in porous media increases with decreasing flow rate, whereas at pH = 7 an electrostatically unfavorable condition has been reported to exist that yields maximum retention [36]. Although several studies of  $TiO_2$  NP transport in porous media have assumed that NP agglomeration in porous media is negligible [17,37], when solution chemistry favors NP-NP interactions over NP-solid matrix interactions,  $TiO_2$  NP agglomeration may be significant [38].

The main aim of this work is to characterize  $TiO_2$  NP suspensions in aqueous solutions, and to investigate the transport of  $TiO_2$  NPs with relatively low concentrations in water saturated porous media in order to improve our understanding of how agglomeration or equivalently particle size can influence NP accumulation in porous media.

# 2. Materials and methods

#### 2.1. Preparation of TiO<sub>2</sub> NPs solutions

The solution pH is known to influence the NP aggregate formation [18,39]. Therefore,  $TiO_2$  NP suspensions at various pH values were prepared using  $TiO_2$  anatase (Aldrich 637254-50G, size < 25 nm), and  $TiO_2$  anatase-rutile (Aldrich, code size < 100 nm). Two different preparation methods were applied. In the first method (M1), the  $TiO_2$  NPs were first dispersed in deionized water (dH<sub>2</sub>O) and then the solution pH was adjusted to the desired value. In the second method (M2), the pH of dH<sub>2</sub>O was adjusted to the desired value, and then the NPs were added to the solution. In both of the preparation methods employed in this work (M1 and M2), the solution pH values were either reduced with HCl 0.1 M or increased with NaOH 0.6 M.

Sonication of NP suspensions may decrease the size of the aggregates and may yield a more stable suspension [39–42]. However, sonication is more appropriate for larger aggregates [41], and the duration of sonication should be carefully selected because excessive sonication may cause re-agglomeration of smaller particles [39,43]. In this study, TiO<sub>2</sub> suspensions were sonicated in an ultrasound bath (Elma, TI-H-5) for various time periods.

#### 2.2. Packed columns

Flowthrough transport experiments were conducted in glass columns with 2.5-cm inner diameter and 30-cm length, which were placed horizontally to avoid gravity effects [30]. The columns were packed with glass spheres of diameter  $d_c = 2$  mm. Each NP solution was injected into the packed column with a peristaltic pump (Masterflex, Cole Parmer). After the end of each experiment the glass spheres were cleaned carefully following the procedure suggested by Bergendahl and Grasso [44]. Briefly, the cleaning procedure consisted of washing the glass spheres with acetone, hexane and concentrated HCl, then washing the glass spheres several times with dH<sub>2</sub>O, soaking them in 0.1 M NaOH solution, re-washing them several times with dH<sub>2</sub>O, and drying them overnight in an oven at 105 °C. The glass column was cleaned with dense chromosulfuric acid, washed several times with dH<sub>2</sub>O, and then dried in an oven.

### 2.3. Analytical methods

The size and zeta potential of the NPs were determined with a zeta sizer (Nano ZS90, Malvern Instruments, UK). The various NP concentrations were determined with a fluorescence spectrophotometer (Cary Eclipse, Varian Australia PTY LTD, Australia). The fluorescence spectrophotometer was equipped with a quartz cuvette ( $10 \text{ mm} \times 10 \text{ mm}$ ), and the excitation/emission wavelength was set at 625 nm [21]. It should be noted that a fluorescence spectrophotometer provides accurate measurements for liquid samples containing low concentration of scattered particles, and small particles [21,45,46].

#### 2.4. Mass recovery and accumulation

The mass recovery,  $M_r$  [–], of the NPs injected into the packed column was quantified by the following expression [47]

$$M_r(L) = \frac{m_0(L)}{C_0 t_p} \tag{1}$$

where  $C_0 [M/L^3]$  is the source concentration of NPs suspended in the aqueous phase,  $t_p [t]$  is the broad pulse duration of the injected NPs, L [L] is the porous medium (column) length, and  $m_0 [tM/L^3]$  is the zeroth absolute temporal moment, which describes the mass of NPs in the breakthrough curve:

$$m_0(L) = \int_0^\infty C(L, t) dt \tag{2}$$

where  $C [M/L^3]$  is the concentration of NPs in suspension.

The normalized mass of the TiO<sub>2</sub> NPs injected in the column, which was accumulated within the column at time t,  $M_a$  [-] was calculated using the following equation:

$$M_{a}(L,t) = \frac{1}{C_{0}t_{p}} \int_{0}^{t} [C_{0} - C(L,t)]dt$$
(3)

#### 3. Results and discussion

# 3.1. Effect of pH on TiO<sub>2</sub> NP suspensions

The particle diameters,  $d_p$  [*L*], and electrokinetic zeta potentials,  $\zeta$  [V], of the various TiO<sub>2</sub> NP suspensions prepared in this work, with both M1 and M2 methods, were measured, listed in Table 1, and graphically presented in Fig. 1. The resulting  $d_p$  values were not sensitive to the method of preparing NP suspensions except for anatase–rutile TiO<sub>2</sub> NP suspensions at pH < 4 and pH > 10. Clearly, both methods of preparation yielded relatively similar  $\zeta$  values for both anatase and rutile–anatase TiO<sub>2</sub> NP suspensions. However, the observed  $d_p$  values for anatase TiO<sub>2</sub> NP suspensions were smaller than those of rutile–anatase TiO<sub>2</sub> NP suspensions. Moreover, the  $\zeta$  values of anatase TiO<sub>2</sub> NP suspensions are shown to be essentially unaffected by the solution pH. This observation



**Fig. 1.** Effect of solution pH on (a) particle size and (b) zeta potential of  $C_o = 10 \text{ mg/L}$  anatase (circles) and  $C_o = 10 \text{ mg/L}$  anatase-rutile (squares) TiO<sub>2</sub> NP suspensions, prepared by methods M1 (open symbols) and M2 (filled symbols). Also, showing is the pH<sub>IEP</sub> = 3.5 for anatase-rutile TiO<sub>2</sub> NP suspensions prepared with method M1. Error bars not shown are smaller than the size of the symbol.

is not in agreement with the findings presented by Fazio et al. [16] and Fang et al. [26], probably due to different initial NP concentrations and NP suspension preparation methods employed. The  $d_p$  and  $\zeta$  values of anatase–rutile TiO<sub>2</sub> NP suspensions are shown in Fig. 1 to decrease with increasing pH. It should be noted that the isoelectric point (IEP), which represents the pH where the

Table 1

Zeta potential and particle size measurements.

Preparation method	рН	$\zeta \pm SD (mV)$	$d_p \pm SD (nm)$
TiO <sub>2</sub> anatase			
M1 no sonication $(t_s = 0)$	2.7	$-7.6 \pm 4.7$	$423.4 \pm 90.0$
	7.5	$-7.4 \pm 4.7$	-
	11.3	$-8.6 \pm 2.6$	252.3 ± 51.1
	12	$-10.1 \pm 2.6$	558.7 ± 73.5
M2 no sonication ( $t_s = 0$ )	4	$-10.2 \pm 3.5$	251.7 ± 37.5
	6.3	$-3.8 \pm 3.5$	227.6 ± 105.7
	7.4	$-2.3 \pm 0.4$	$91.0 \pm 0.8$
	10.4	$-9.5 \pm 5.5$	340.14 ± 126.7
TiO <sub>2</sub> anatase–rutile			
M1 no sonication ( $t_s = 0$ )	2.9	$6.9 \pm 1.5$	115.8 ± 26.2
	6.1	$-23.3 \pm 0.85$	745.7 ± 141.8
	11.2	$-39.7 \pm 2.9$	285.5 ± 121.5
M2 no sonication ( $t_s = 0$ )	4	$-12.9 \pm 0.51$	809.7 ± 68.3
	5.8	$-18.4 \pm 0.7$	677.7 ± 127.2
	9	$-28.7 \pm 0.7$	647.0 ± 65.0
	10.9	$-44.2 \pm 1.3$	530.0 ± 57.0
M2 30 min sonication ( $t_s = 30 \text{ min}$ )	4	$-9.8 \pm 0.8$	1054.0 ± 120.4
	5.8	$-29.7 \pm 0.7$	437.9 ± 36.6
	10.9	$-42.3 \pm 0.85$	326.0 ± 38.0
M2 30 min sonication & 1 day aging ( $t_s = 30 \text{ min}, t_a = 1 \text{ d}$ )	4	$-12.0 \pm 1.3$	811.0 ± 260.0
	5.8	$-26.3 \pm 0.5$	417.9 ± 3.4
	10.9	$-41.5 \pm 0.8$	267.0 ± 61.5

electrophoretic mobility changes from positive to negative, for anatase–rutile suspensions prepared by the M1 method, was found to be equal to  $pH_{IEP} = 3.5$  (see Fig. 1b). A similar  $pH_{IEP}$  has also been reported in the literature for a rutile TiO<sub>2</sub> NP suspension [48]. However, depending on the manufacturer,  $pH_{IEP}$  values for TiO<sub>2</sub> NP suspensions can vary considerably [48,49]. The  $\zeta$  values for anatase–rutile TiO<sub>2</sub> NP suspensions prepared by the M2 method, and for anatase TiO<sub>2</sub> NP suspensions prepared by both M1 and M2 methods, were negative for all pH values examined in this study (see Fig. 1b); but the possibility of switching to positive  $\zeta$  values at lower pH values cannot be ruled out.

Relatively small, suspended particles, with high absolute  $\zeta$  values are expected to exhibit better stability than suspended particles with low absolute  $\zeta$  values. At low absolute  $\zeta$  values, aggregates are formed because attracting forces between suspended particles are stronger than repulsive forces [26,50–52]. For the pH range examined in this work, anatase TiO<sub>2</sub> NP suspensions, formed smaller aggregates than anatase–rutile TiO<sub>2</sub> NP suspensions; however, the anatase TiO<sub>2</sub> NP suspensions exhibited smaller absolute  $\zeta$  values than anatase–rutile TiO<sub>2</sub> NP suspensions.

#### 3.2. Effect of sonication time

Sonication of suspended particles is known to reduce the size of aggregates, and also to provide relatively stable suspensions [39–42]. To obtain the optimum sonication period,  $t_s$  [t], several concentrations of anatase–rutile NP suspensions, prepared by method M2 at pH = 7 were sonicated for varying time periods, as shown in Figs. 2a and 3. Clearly, the results suggest that the mean  $d_p$  of NP suspensions without sonication ( $t_s = 0$ ) and with sonication ( $t_s > 0$ ) increases with increasing initial NP concentration. This result is consistent with the observations made by Brunelli et al. [21] who reported that the initial TiO<sub>2</sub> NP concentration plays a



**Fig. 2.** Effect of sonication time on: (a) particle size, and (b) zeta potential of anatase-rutile  $TiO_2$  NP suspensions prepared with method M2, as a function of initial NP concentration. Error bars not shown are smaller than the size of the symbol.



**Fig. 3.** Effect of sonication time on particle size of anatase–rutile  $TiO_2$  NP suspensions prepared with method M2, at two relatively high initial concentrations. Error bars not shown are smaller than the size of the symbol.

key role on NP agglomeration and sedimentation in aquatic solutions, whereas other parameters (e.g., ionic strength, salt content and dissolved organic carbon) are of less importance. However, it should be noted, for relatively small initial concentrations  $(C_0 \leq 5 \text{ mg/L})$ , the range of  $d_p$  values observed for  $t_s = 2 \text{ h}$  is much greater than that for  $t_s = 1$  h and 3 h (see Fig. 2a). For relatively high initial concentrations ( $C_o \ge 500 \text{ mg/L}$ ), the time period of sonication plays an important role because the mean  $d_p$ , as well as its range decreases with increasing  $t_s$  (see Fig. 3). These observations are in agreement with the results presented by Horst et al. [39] and Jiang et al. [41] who noted that sonication is appropriate for the breakage of large agglomerates, but it is not effective for the small sizes, because extended sonication may cause re-agglomeration due to increased collision frequency of smaller particles. The data presented in Fig. 2b suggest that there is no clear trend between the measured  $\zeta$  values and t<sub>s</sub> as a function of  $C_o$ .

Anatase–rutile TiO<sub>2</sub> NP suspensions with  $C_o = 10$  mg/L, prepared by method M2 at various pH values, were also sonicated for  $t_s = 30$  min. The  $d_p$  and  $\zeta$  values of the NP suspensions were measured immediately after sonication, and once again after a time delay or aging time,  $t_a$  [t], of one day ( $t_a = 1$  d). The experimental results are graphically illustrated in Fig. 4, and suggest that with or without sonication and aging,  $d_p$  and  $\zeta$  decreased with increasing solution pH. Worthy to note is that at pH > 6,  $d_p$  decreased with sonication compared to the case of no sonication, and decreased more with aging (see Fig. 4a). Also, at pH ~ 6,  $\zeta$  became more negative (more stable) with sonication, and even more negative with aging; whereas, at pH ~ 6, sonication and aging did not significantly affect  $\zeta$ . Consequently, a short time of sonication ( $t_s \sim 30$  min) at pH > 6 are considered optimum conditions for anatase–rutile TiO<sub>2</sub> NP suspensions.

#### 3.3. Transport experiments

All transport experiments were conducted with anatase TiO<sub>2</sub> NP suspensions prepared by method M2 at pH = 7, without sonication and aging ( $t_a = t_s = 0$ ). Note that at pH = 7, anatase TiO<sub>2</sub> NP suspensions exhibit relatively small initial  $d_p$  (see Fig. 1a) and thus straining may not be significant, but the measured  $\zeta$  values at pH = 7 suggest that anatase TiO<sub>2</sub> NP suspensions are more prone to agglomeration than anatase–rutile TiO<sub>2</sub> NP suspensions (see Fig. 1b). Three different concentrations of anatase TiO<sub>2</sub> NP suspensions were used in this study:  $C_o = 2.5$ , 5, and 7 mg/L. For each transport experiment a TiO<sub>2</sub> NP suspension solution was injected into the packed column with a flow rate of Q = 2 mL/min. When the effluent NP concentration reached a plateau, the TiO<sub>2</sub> NP suspension solution was replaced with dH<sub>2</sub>O. The injection period,  $t_p$  [t], of the NP suspension was recorded and listed in Table 2. Effluent



**Fig. 4.** Effect of solution pH on (a) particle size and (b) zeta potential of  $C_o = 10 \text{ mg/L}$  anatase–rutile TiO<sub>2</sub> suspensions, prepared by method M2 without sonication ( $t_s = 0$ , squares), with  $t_s = 30 \text{ min}$  (diamonds), and with  $t_s = 30 \text{ min}$  followed by one day aging ( $t_a = 1 \text{ d}$ , triangles). Error bars not shown are smaller than the size of the symbol.

ent liquid samples were collected periodically and analyzed for TiO<sub>2</sub> NP concentration and  $d_p$ . The breakthrough data collected from the various experiments were presented in Fig. 5. Also, the normalized maximum (peak) NP concentration from each breakthrough data set was graphically presented in Fig. 6. Note that experiments 1–3 were conducted with  $C_o = 2.5 \text{ mg/L}$  (see Fig. 5a), experiments 4–6 with  $C_o = 5 \text{ mg/L}$  (see Fig. 5b), and experiments 7 and 8 with  $C_o = 7 \text{ mg/L}$  (see Fig. 5c). Clearly, breakthrough data obtained from experiments conducted with identical  $C_o$  are somewhat different, especially for  $C_o = 5 \text{ mg/L}$  (see Fig. 5b). This variability is attributed to randomly occurring NP agglomeration within the packed column. However, as expected, the observed peak concentrations of the various transport experiments exhibited a trend of increasing values with increasing  $C_o$  (see Fig. 6).

The measured effluent  $d_p$  values from all of the transport experiments conducted in this study are presented in Fig. 7, together with the corresponding initial  $d_p$  values of the injected NP suspensions. The effluent  $d_p$  values (see Fig. 7), as well as the mean effluent  $\langle d_p \rangle$  values (see Table 2) were significantly different than the

Table 2				
Experimental	conditions	and	estimated	$M_r$ values.



**Fig. 5.** Normalized effluent anatase  $TiO_2$  NP concentration data for the transport experiments conducted with initial concentration: (a)  $C_o = 2.5 \text{ mg/L}$ , (b)  $C_o = 5 \text{ mg/L}$ , and (c)  $C_o = 7 \text{ mg/L}$ . Here Q = 2 mL/min.

initial  $d_p$  values of the injected NP suspensions. The observed variability in the measured effluent  $d_p$  values (see Fig. 7) suggests that substantial NP agglomeration took place within the packed column during the transport experiments. This observation is in agreement with earlier work on NP agglomeration in porous media [38]. Note that there is no particular trend in the measured effluent  $d_p$  values, because there is randomness in the NP agglomeration process within the packed column. Also, effluent  $d_p$  values do not provide

Exp. No.	$C_o (mg/L)$	Q (mL/min)	$t_p$ (min)	Initial $d_p$ (nm)	Mean effluent $\langle d_p \rangle$ (nm)	$M_r^a$
1	2.5	2	85	179.0	190.2	0.13
2	2.5	2	95	278.0	182.0	0.16
3	2.5	2	105	154.5	168.1	0.19
4	5	2	120	207.4	156.6	0.25
5	5	2	83	240.0	273.3	0.17
6	5	2	105	92.0	166.0	0.46
7	7	2	100	51.3	140.9	0.36
8	7	2	120	138.9	86.4	0.32

<sup>a</sup> Evaluated with (1).



Fig. 6. Normalized maximum effluent anatase  $TiO_2$  NP concentration as a function of the injected initial  $TiO_2$  NP concentration.

a true representation of  $\langle d_p \rangle$  for the agglomerates formed and retained within the packed column.

The normalized mass accumulated within the packed column,  $M_a$  [-], during each of the transport experiments conducted in this study was determined with Eq. (3), and presented graphically as a function of time in Fig. 8. The estimated normalized accumulated mass curve for each transport experiment was increasing linearly with time for a time period equal to  $t_p$ , and then it asymptotically



**Fig. 7.** Observed effluent NP sizes from transport experiments with: (a)  $C_o = 2.5 \text{ mg/}L$  TiO<sub>2</sub>, (b)  $C_o = 5 \text{ mg/}L$  TiO<sub>2</sub>, and (c)  $C_o = 7 \text{ mg/}L$  TiO<sub>2</sub>. Here Q = 2 mL/min. The dashed lines represent the injected initial NP sizes of the various transport experiments. Error bars not shown are smaller than the size of the symbol.



Fig. 8. Accumulated mass of anatase  $TiO_2$  retained in the packed column of each experiment as a function of time.

reached a plateau. The time to reach the asymptotic plateau from the maximum  $M_a$  value of each experiment was affected by NP detachment from the column solid matrix, which, as expected was more pronounced for smaller NPs. Clearly, in experiments 6– 8, which were conducted with the smallest initial  $d_p$  values, more time was required for  $M_a$  to reach its asymptotic value.

The mass recovered from each of the transport experiments conducted in this study was determined with Eq. (1), listed in Table 2, and graphically illustrated in Fig. 9a, in conjunction with the corresponding initial  $d_p$  concentrations shown in Fig. 9b. Clearly, the mass recovery of the injected TiO<sub>2</sub> NPs was increasing with increasing  $C_o$ . This result is intuitive because as shown in Fig. 9b, the  $d_p$  of the injected NP suspensions was decreasing with increasing  $C_o$ . Assuming that possible agglomeration of the injected anatase TiO<sub>2</sub> NPs within the column at pH = 7 would yield particles with  $d_p < 300$  nm (see Fig. 1a), then straining and wedging should



Fig. 9. Effect of initial concentration on: (a) mass recovery, as related to (b) the initial size of the injected anatase  $TiO_2$  NPs.

be ruled out of possible mechanisms of mass accumulation within the packed columns examined in this study, because the corresponding colloid to collector diameter ratios ( $d_p/d_c < 0.00015$ ) were well below the suggested threshold of 0.004 [53] or 0.003 [54]. Consequently, the observed accumulation of TiO<sub>2</sub> NPs within the column should be attributed only to physicochemical attachment processes.

#### 4. Conclusions

The experimental results of this study revealed that smaller aggregates with smaller absolute  $\zeta$  values were formed with anatase TiO<sub>2</sub> NP suspensions than anatase-rutile TiO<sub>2</sub> NP suspensions. The method of preparing NP suspensions did not significantly affect the resulting  $\zeta$  values, but affected the resulting  $d_n$  values for anatase-rutile TiO<sub>2</sub> NP suspensions at pH < 4 and pH > 10. Sonication decreased the aggregate size for NP suspensions of relatively high initial concentrations, but extensive sonication could lead to re-agglomeration of low NP concentrations with small particles. Also, sonication and aging reduced NP aggregate size at relatively high pH values. The results from the transport experiments with anatase TiO<sub>2</sub> NP suspensions suggested that mass recovery of the injected TiO<sub>2</sub> NPs could increase with increasing C<sub>o</sub> and decreasing  $d_p$ . Furthermore, NP accumulation within the packed columns was significant due to agglomeration in conjunction with physicochemical attachment rather than straining or wedging.

#### Acknowledgments

This research was partially funded by the European Union (European Social Fund-ESF) and Greek National Funds through the Operational program "Education and Lifelong Learning" under the action Aristeia I (Code No. 1185). The authors are thankful for the laboratory assistance provided by D. Vassilopoulos, and Dr. I.D. Manariotis. This is collaborative work between members of the BioMet network.

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